



# Conservation from the inside-out: Winning space and a place for wildlife in working landscapes

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## Abstract

1. Protected areas fall far short of securing the space needed to sustain biodiversity and ecosystem function at a global scale and in the face of climate change.
2. The prospects of conserving biodiversity in working landscapes help buffer the insularization effects of protected areas and hold great potential for biodiversity conservation on a landscape scale but depend on finding adequate space and a meaningful place in the lives of rural land users.
3. Using a case study in southern Kenya, we show that the conservation of large open landscapes, biodiversity and the coexistence between wildlife and livestock can be achieved indirectly by reinforcing pastoral practices that depend on open space, mobility, social networks and institutional arrangements governing common-pool resources.
4. Pastoral practices and wildlife both depend on large multiscale interactions within interlinked social and ecological systems, which are threatened by land fragmentation, alienation and degradation.
5. We show that large open spaces can be maintained by using a conservation approach starting from within community aspirations that emphasize the links between livelihoods, productivity, efficiency and resilience in pastoral economies and the secondary benefits of wildlife enterprises.
6. Scaling up from an ecosystem to multi-scale approach benefits pastoral communities by building resilience and new economic opportunities. In the process, the expanded scale conserves regional biodiversity and large free-ranging herbivore and carnivore populations underpinning ecosystem function and the nationally important tourism industry centered on the Kenya–Tanzania boundary.
7. The ‘inside-out’ approach to the conservation of wildlife and biodiversity is place-based, draws on local knowledge and informal governance arrangements and avoids the stigma of wildlife conservation driven by outside agencies.

The art of land doctoring is being practiced with vigor, but the science of land health is yet to be born. Aldo Leopold: A Sand County Almanac.

We protected wildlife from hunters and wildlife protected us from drought. Coexistence is the essence of survival for us both. Maasai elder noting why traditional livestock practices conserve wildlife.

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8. The human-centered approach reinforces land health and spatial connectivity and encourages multi-level and distributed governance arrangements embedded in large regional and national jurisdictions.

#### KEYWORDS

coexistence, community-based conservation, conservation governance, landscape conservation, natural resource management, pastoralism, wildlife, working landscapes

## 1 | THE EVOLUTION OF CONSERVATION

The scope of conservation has expanded continuously in the modern era, from natural resource management in the 18th century, to regulated hunting and protected area set-asides in the 19th century, and the conservation of species, biotic communities, ecosystems and biodiversity in the 20th century (Watson, Dudley, Segan, & Hockings, 2014). The expanded scope reflects changing human sensibilities, views and uses of nature (Nash, 1989; Thomas, 1983) and, more recently, the aspirations of modern states and the international community (Adams, 2005; Sitarz, 1993; United Nations, 1992; Watson et al., 2014; Western & Pearl, 1989; Western, Waithaka, & Kamanga, 2015; WWF, 1980). The philosophical and ethical foundations of conservation have in turn deepened from a largely utilitarian creed to encompass recreational, romantic, spiritual, educational, scientific, intrinsic and other values (Callicott, 1990; Meine, Soulé, & Noss, 2006; Vucetich, Bruskotter, & Nelson, 2015).

Changes in societal perceptions are echoed by the global perspective of the World Parks Congress (WPC) over the last half-century. The vision for national parks at the 1972 WPC projected the traditional view of areas set aside to protect natural wonders and wildlife for human recreation and enjoyment. The vision expanded to include parks for sustainable development at the 1982 WPC, and human well-being at the 2014 congress (McNeely, 1993; WPC, 2014). In the process six categories of internationally accredited protected areas have been recognized, ranging from Category I: strict nature reserves for scientific and wilderness protection, to Category V: land and seascapes where human uses have produced areas of distinctive aesthetic, ecological and cultural value, and Category VI: areas managed mainly for the sustainable use of natural ecosystems (Dudley & Stolton, 2008).

Although the WPC has raised the target for terrestrial coverage from 10% in the 1980s to 20% for 2020 (WPC, 2014), the coverage falls far short of the area needed to sustain biodiversity for several reasons. First, historically, most protected areas were set aside for scenic, recreational and aesthetic reasons and for specific wildlife attractions, rather than for biodiversity or conserving ecosystem viability (Bennett et al., 2009; Fynn & Bonyongo, 2011; Jenkins, Pimm, & Joppa, 2013; Western & Gichohi, 1993). Second, most protected areas were set aside in lands marginal for development, not to conserve biodiversity (Joppa & Pfaff, 2009; Pressey & Bottrill, 2008; Venter et al., 2017). Third, most protected areas are too small to avoid a loss of species due to insularization, habitat fragmentation

and ecological disruption (Newmark, 2008). Fourth, over one-third of land user protected areas is degrading due to human pressures (Jones et al., 2018). Finally, the prospects of retrofitting protected areas to maximize biodiversity coverage are limited, given existing land tenure and uses (Pressey, Visconti, & Ferraro, 2015). Calls for allocating half the earth to protected areas, primarily free of human activity (Dinerstein et al., 2017; Wilson, 2016), face insurmountable political obstacles and fail to redress the systemic shortcomings of existing protected areas (Büscher & Brockington, 2017). The alternative is to consider finding space in lands used for human uses compatible with conserving biodiversity and maintaining the ecological health of the land.

## 2 | CONSERVATION IN THE HUMAN REALM

The prospects of conserving biodiversity within the human realm by sustaining the health of the land and its capacity for renewal has long been recognized as an evolutionary possibility and ecological necessity (Leopold, 1949). Leopold, in calling for a land ethic, noted that the art of land doctoring is being practiced with vigour, but the science of land health is yet to be born. The land ethic is gaining recognition as being essential for halting global losses of biodiversity (Kremen & Merenlender, 2018). The principles of the World Conservation Strategy (WWF, 1980), Agenda 21 (Sitarz, 1993), the Convention on Biological Diversity (United Nations, 1992) and the Millennium Ecosystem Assessment (Assessment, 2003) all underscore the importance of land health for sustainable development. Yet, land health also depends on cultural values and governance institutions (Brockington et al., 2019; Ostrom, 2009). Many traditional societies, referred to as biocultures by Nicolay Vavilov (Nabhan, 2012), evolved husbandry practices and cultures that have sustained the health of the land for generations in the face of environmental perturbations and climate change.

Conserving biodiversity in landscapes transformed by farming, ranching and other human uses depends on wildlife having access to their land and in their lives on terms favourable to them. The options for doing so include land sharing, land saving and land sparing (Fischer et al., 2014). The mixed-use and implicit coexistence in land sharing move biodiversity conservation beyond the sharp distinction between natural and human-modified landscapes implicit in protected areas to a land sparing-sharing continuum based

on the intensity of uses (Phalan, 2018; Western & Pearl, 1989). The scale ranges from heavily transformed and intensively used lands such as monoculture farmlands with little biodiversity to extensive uses in less modified landscapes such as grass-fed ranching, agroecological farming, renewable forestry practices, recreational uses of wildlands rich in biodiversity and extensive mobile pastoralism (Kremen & Merenlender, 2018). Land sparing-sharing models can contribute significantly to biodiversity conservation, though not necessarily from a livelihood and social perspective (Phalan, 2018).

The expansion of nature conservation from the protection of habitats and species to sustaining biodiversity launched by the World Conservation Strategy in 1980 calls for large-scale planning across a broad range of land uses, users, jurisdictions and agencies (WWF, 1980). Biodiversity conservation in the human realm, especially at the scale needed to conserve species in the face of projected climate and land use changes in the 21st century (Newbold, 2018), must also fit in with other land uses and minimize or offset the socio-economic losses incurred (Donaldson, Wilson, & Maclean, 2017).

If scaling up biodiversity conservation has great potential, it also faces severe obstacles (Curtin, 2015; Western & Pearl, 1989). Community-based conservation, largely focused on ecosystem-level conservation, has shown some success in conserving wildlife and biodiversity in rural landscapes and in improving socio-economic development (Naidoo et al., 2016; Oldekop, Holmes, Harris, & Evans, 2016; Shahabuddin & Rao, 2010). Scaling up biodiversity conservation to a large landscape level calls for yet wider networks and jurisdictions cutting across varied land uses, user interests and jurisdictional boundaries (Arts et al., 2017; Sayer et al., 2013; Scarlett & McKinney, 2016).

Several such landscape approaches have emerged over the last few decades, ranging from the Yellowstone to Yukon Conservation Initiative (Merrill, 2005), Colorado River Initiative (Adler, 2007), Malpai Borderlands Group (Curtin, 2002), Kavango Zambezi Transfrontier Conservation Area (Cumming, 2008) and the Kenya-Tanzania Borderlands Conservation Initiative (BCI).<sup>1</sup> In many cases, the scale needed to secure sufficient space for biodiversity to accommodate large-scale migrations of species traverses several national boundaries. Many of these initiatives have struggled to reach their landscape-scale goals due to a limitation in economic resources for biodiversity conservation.

The tools for encouraging landscape biodiversity conservation are primarily economic in nature and include (a) conservation leases and easements, (b) payment for ecosystem services, (c) cost offsets, and (d) consumptive and non-consumptive uses of wildlife (Bedelian & Ogutu, 2017; Naidoo et al., 2016; Nelson et al., 2010). New economic accounting methods for valuing natural capital and ecological services, such the Total Economic Value (Costanza et al., 1997) and the System of Environmental-Economic Accounts (Jasch, 2003), emphasize the multiple benefits of conserving biodiversity. These tools complement the efforts of parks to expand

the reach of nature conservation and alleviate human threats to biodiversity.

Large herbivores and carnivores have proven especially hard to conserve outside protected areas given a history of overhunting, displacement, the large spaces and intact habitats they need to survive and the dangers and competition they pose to people (Tilman et al., 2017). Smith, Smith, Lyons, and Payne (2018) show that a strong down-sizing of species has characterized the human impact on ecosystem structure and function over at least the past 125,000 years and is likely to continue to do so far into the future.

Despite these obstacles, there is considerable scope for conserving large herbivores and carnivores in the pastoral areas of Africa and rangelands around the world, far more than in protected areas (Western & Pearl, 1989). The East African savannas reflect the long co-evolution of humans and wildlife that survived both the Pleistocene megafaunal extinctions and colonial-era decimation of wildlife in other places (Adams & McShane, 1992). The reasons are due in part to the cultural and use values of wildlife (Western, 2019), the seasonal mobility of pastoralists and the milk-based pastoralist economy. If classified by the IUCN criteria, the pastoral-dominated savannas would be recognized as the equivalent of Category V landscapes where human uses have produced areas of distinctive aesthetic, ecological and cultural values, and Category VI areas managed mainly for the sustainable use of natural ecosystems.

In this article, we focus on wildlife and biodiversity conservation in the East African savannas but also draw broader inferences for the rangelands which cover 25% of the earth's surface and working landscapes more generally. Successfully conserving wildlife in the savannas depends on how herding people manage their lands (Groom & Western, 2013) and finding places for wildlife which fit within the livelihood needs and cultural values of communities (de Pinho & Ellis, 2009). We argue that the space and mobility for sustaining large mammals can be secured indirectly through an approach we term conservation from the 'inside-out'. By inside-out we mean drawing on husbandry and conservation practices used to maintain the productivity and resilience of pastoralism or other land uses that directly or indirectly maintain large free-ranging wildlife movements in the process. Though overlapping, inside-out conservation is distinguished from bottom-up approaches to wildlife conservation in using primary livelihood considerations to win space for wildlife indirectly rather than through direct incentive-based approaches.

Ultimately, the inside-out approach aligns biodiversity conservation to land health and sustainable husbandry practices, thus widening the scope for nature conservation to the majority of the earth's surface without necessarily relying on large capital investments or tackling biodiversity as the primary focus. Here, we use a case study that looks at the adaptation of traditional pastoral practices in East Africa to the fast-changing global age, and by doing so offers prospects for sustaining land health, improving livelihoods and conserving biodiversity.

### 3 | PASTORAL AND WILDLIFE LANDSCAPES

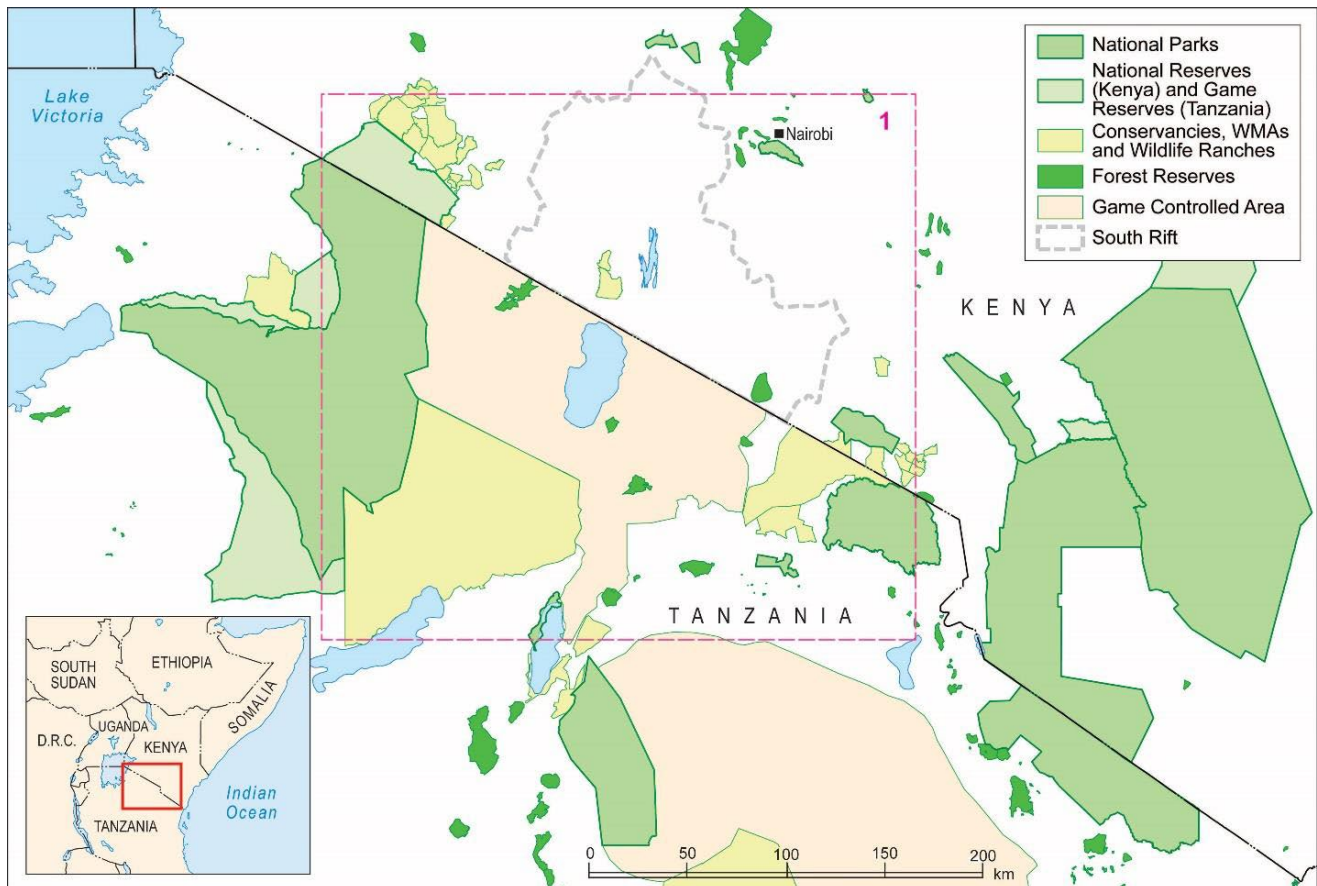
Maasai pastoralists occupy much of the 100,000 km<sup>2</sup> area that spans the Kenya–Tanzania borderlands and the Rift Valley (Figure 1). In this region, bimodal rainfall varies from 250 to 1,700 mm across altitudes ranging from 600 m on the floor of the Rift Valley to over 5,000 m on Kilimanjaro. The range of biomes cutting across the rainfall and altitudinal gradients make the Kenya–Tanzania borderlands one of the richest biodiversity regions in Africa and mammalian diversity centres on Earth (Jenkins et al., 2013). The region also has some 16 national parks and reserves stretching from Serengeti and Maasai Mara in the West to Tsavo and Mkomazi in the East (Figure 1). The continuous patchwork of grasslands, bushlands and woodlands reaching across this region also support the greatest abundance of pastoral livestock and wildlife in Eastern Africa (Ogotu et al., 2016).

The Maasai people are made up of 13 politically semi-autonomous sections, each sharing the same clan and age-set groupings. The overlapping clans and age-sets move seasonally between wet and dry season grazing areas, largely in synchrony with wildlife migrations

(Western & Nightingale, 2005) by using reciprocal social ties that span the seasonal ranges and foster connections to adjoining Maasai sections (Spear & Waller, 1993).

The seasonally coordinated movements between wet and dry season ranges define the community of users and sustain their livestock herds except in extreme years when herders often move to distant refuges to evade droughts. Drought refuges are typically in the wetter regions of the landscape, such as highland pastures, wetlands and areas receiving recent rains (Fynn, Murray-Hudson, Dhiwayo, & Scholte, 2015). The drought movements expand the scale of use from the annual range of a few thousand square kilometres to tens of thousands, negotiated through their reciprocal arrangements (Western & Finch, 1986).

The conservation of the grazing commons to ensure herd productivity and resilience is deeply rooted in Maasai governance and herding practices (Spear & Waller, 1993). A family's survival and well-being are bound to the welfare of its livestock and the availability of pasture and water. The Maasai nevertheless have no word for conservation. The link between rainfall, pasture production, herd productivity, family welfare and the maintenance of commons resources is, instead, incorporated in the concept of *erematere*. Akin



**FIGURE 1** Map of the conservation and protected areas network in the Kenya–Tanzania borderland, highlighting the location of our case study within the region. Our case study looks at three scales: Shompole and Olkiramatian conservancies in the centre of the map, the South Rift at the landscape level and the regional level across the whole map. The red box indicates the extent of Figure 3. Data are from the World Database of Protected Areas ([www.protectedplanet.org](http://www.protectedplanet.org)), South Rift Association of Landowners ([www.soralo.org](http://www.soralo.org)), Maasai Mara Wildlife Conservancies Association ([www.maraconservancies.org](http://www.maraconservancies.org)) and BigLife Foundation ([www.biglife.org](http://www.biglife.org))

to Vavilov's biocultures and Leopold's land ethic, *eramatare* is an ethos embedded in husbandry practices, cultural customs and the governance of Maasai society. *Eramatare* linkages stretch across the landscape through social networks, giving families access to the resources needed to sustain them through the seasons and in times of drought. *Eramatare* also extends to wildlife, which holds many values and uses among the Maasai, including food, clothing, medicine, ornamentation, utensils, clan symbols and aesthetic appeal (Western, 2019). Many species, including eland and buffalo, are regarded as 'second cattle' and are used as a standby food source in times of drought (Western, 1997).

## 4 | HUMAN-DRIVEN CHANGES IN EAST AFRICA

Starting in the 1940s, concerns over the impact of population growth, land transformation and poaching of wildlife led to the creation of national parks and reserves presently covering 15% of Tanzania and 8% of Kenya. By the 1970s recognition that national parks are too small to avoid extinctions and insufficient to protect the wildlife migratory routes led to policies for engaging communities in wildlife conservation through tourism and hunting revenues (Western & Pearl, 1989). In the ensuing decades the policy shift resulted in an expansion of wildlife conservation onto community lands by encouraging the creation of wildlife conservancies in Kenya and Wildlife Management Areas in Tanzania (KWCA, 2016).

Despite the spread of community-based conservation in Africa's rangelands, its success hinges on the future of pastoral economies and cultures. Pastoralism, the dominant form of land use in the sub-arable savannas, accounts for over 90% of the large mammal biomass in Kenya (Ogutu et al., 2016). Despite the economic potential of tourism and sport hunting, wildlife revenues contribute to only a small portion of household income (Kristjanson & Trench, 2009). The strong cultural identity and social bonds rooted in livestock also make it likely that pastoralism will remain the primary form of livelihood even where wildlife revenues are a significant portion of family incomes.

Pastoralism nevertheless faces many of the same threats as wildlife, namely the loss of space and mobility; land use changes and land degradation; a loss of livestock production; and decreasing resilience to droughts (Boone, 2005; Hobbs et al., 2008). The loss of ecological adaptability to environmental perturbations is further compounded by a breakdown in the traditional governance institutions rooted in social reciprocity which regulate pasture use and minimize local risks of drought, disease and other hazards (Mwangi & Ostrom, 2009). The breakdown stems from several factors arising from pre- and post-colonial government policies, including government-mandated regulations of the rangelands, the creation of wildlife parks and reserves, forced eviction of pastoral communities for agricultural development, sedentarization and subdivision policies for rangelands and the replacing and breakdown of traditional decision-making practices (Mwangi &

Ostrom, 2009). Because of these common threats, redressing the threats to pastoralism and pastoral lands indirectly alleviates the same threats to wildlife. Further, conserving the cultural and governance practices underpinning the mobility and resilience of pastoral herds also reduces human-wildlife conflict and the prospects for continued coexistence (Western, 2019).

Here we look at an example of how two adjacent Maasai communities in Kenya's South Rift (Figure 1) are drawing on both traditional and contemporary knowledge of husbandry and governance practices to sustain pastoralism and conserve wildlife. We then look at how the expansion of such ecosystem-level efforts can benefit from increased scales of governance and management to not only build livestock production and resilience but scale up to large landscapes to conserve biodiversity.

## 5 | CONSERVATION FROM THE INSIDE-OUT

### 5.1 | The benefits of landscape scale to pastoralists and wildlife

Among pastoral communities free-ranging movements give livestock access to resources over large regions where forage and water sources are patchy and ephemeral. Mobility allows herders to track the richest pastures through the season and minimizes exposure to drought, disease, local pasture degradation and perturbations in semi-arid areas where rainfall is highly variable in time and space (Ash, Gross, & Smith, 2004; Boone, 2005; Hobbs et al., 2008; Wang et al., 2006; Western, 1982; Western & Finch, 1986). These ecological benefits of large-scale mobility in areas where resource availability is highly stochastic are reflected in the energy bonus of improved digestive efficiency, growth rates and milk yields for pastoral livestock and wildlife alike (Illius & O'Connor, 2000; Owen-Smith, 2004; Wang et al., 2006). In the case of pastoralists, the scale of use is tied to the scale of social networks allowing free and safe movement (Western, 2019). Large social networks have in turn been shown to increase the benefit to individuals through a 'return to scale' across a range of societies from hunter-gatherers to agrarian and urban economies (Hamilton, Milne, Walker, & Brown, 2007; West, 2017).

The ecological benefits of mobility in offsetting stochastic resource availability apply widely regardless of habitat and landscape variability. The benefits become far more beneficial, however, in landscapes varying in elevation, climate, soils, hydrology, habitat and plant nutrition. In such varied landscapes, the functional heterogeneity at scale adds the productivity, diversity and resilience of large herbivores (Fynn, Augustine, Peel, & Garine-Wichatitsky, 2016; Owen-Smith, 2002, 2004).

In the case of the African savannas, wildlife survival, abundance and resilience to seasonal flux and drought are also scale-dependent, and like livestock, they depend on the ecological benefits accruing across large functionally heterogeneous landscapes (Fryxell et al., 2005; Owen-Smith, 2004; Western & Gichohi, 1993). Species such

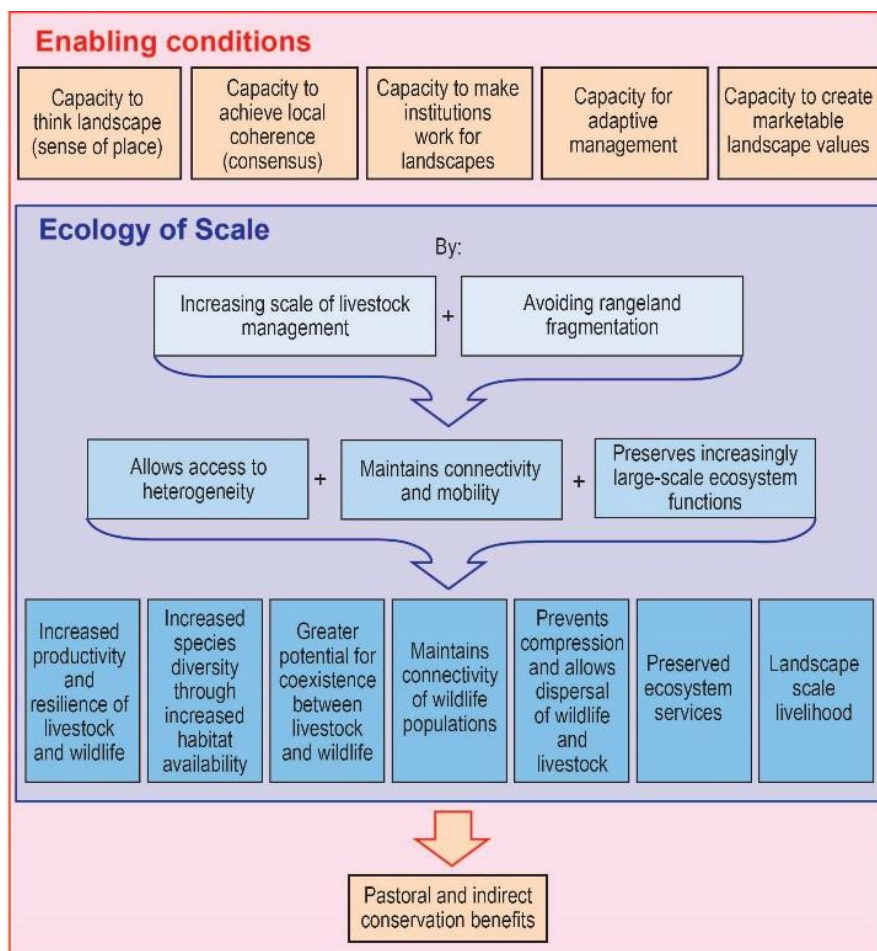
as elephants, lions, wild dogs, giraffe and migratory wildebeest, zebra and gazelle in the Kenya–Tanzania borderlands cover thousands of square kilometres in the course of seasonal movements (Dolrenry, Stenglein, Hazzah, Lutz, & Frank, 2014; Fryxell et al., 2005; Mose, Nguyen-Huu, Western, Auger, & Nyandwi, 2013; Osipova et al., 2018). In that species richness, habitat diversity and ecosystem integrity all increase with landscape heterogeneity (Peterson, Allen, & Holling, 1998; Figure 3), conserving metapopulations of landscape species—species using a large geographic area which includes a wide variety of other species—conserves biological diversity and integrity in the process. Large open landscapes also avoid the compression effects of large mammals on biodiversity (Western & Maitumo, 2004) and rangeland health (Hobbs et al., 2008; Western, Mose, Worden, & Maitumo, 2015) through seasonal movements and population dispersal (Mose & Western, 2015).

Scale and mobility in increasing functional heterogeneity also facilitate coexistence between livestock and wildlife by increasing foraging options, minimizing competition (Fynn et al., 2016; Tyrrell, Russell, & Western, 2017) and expanding the scope for the spatial and temporal separation of wildlife and livestock. Temporal separation in turn reduces pathogen transmission, crop-raiding and livestock depredation (Valls-Fox et al., 2018; Western, 2019).

Finally, scale and mobility increase the structural and functional heterogeneity of landscapes through disturbance effects created

by the differential grazing and browsing impacts (Fuhlendorf et al., 2016; Fynn et al., 2016) important to conserving the richness and integrity of ecosystems. Examples of patch dynamics caused by the seasonal movements of livestock and wildlife to optimize foraging include the interaction of elephants and livestock in creating a shifting mosaic of habitats (Asner, Vaughn, Smit, & Levick, 2016), the intensity, timing and scale of fires influencing biodiversity and forage quality (Fynn et al., 2016; Morrison et al., 2018) and abandoned settlements in creating nutrient hotspots and habitat succession (Muchiru, Western, & Reid, 2009; Vuorio, Muchiru, Reid, & Ogutu, 2014). The expanded scope for coexistence linked to scale also favours ecotourism, hunting and recreation activities that diversify pastoral economies and sustains ecological services such as carbon sequestration, nutrient cycling, and water supplies at local, national and global scales (Yahdjian, Sala, & Havstad, 2015).

Expanding the scale of landscape management from an inside-out approach is driven by the vested interests of pastoralists and agro-pastoralists in sustaining their livelihood productivity and resilience and, incidentally, avoids the negative impacts of fragmentation of rangelands (Groom & Western, 2013; Hobbs et al., 2008). Conservation approached from the self-interests of herders and their social networks shift the focus from top-down outside-driven programs to culturally embedded and community-based approaches to conservation (Figure 2).



**FIGURE 2** A diagrammatic representation of conservation from the inside-out for rangeland and pastoral systems. Successful resource governance of the pastoral rangelands depends on enabling conditions (Arts et al., 2017) linked to the ecology of scale and social networks (Cumming, Olsson, Chapin, & Holling, 2013). Scale benefits pastoralists through increased mobility, herd production, resilience and the maintenance of land health and, in the process, conserves biodiversity and ecosystem services. The benefits and implications of the ecology of scale can act across multiple scales (ecosystem, landscape, region)

Social-ecological system theory (SES) and landscape governance theory recognize that successful management of common property resources depends on managing the commons through social networks and ruled-based institutions with coherence at the appropriate social and ecological scales, the capacity for adaptive management through devolved management rights and marketable landscapes (Figure 2; Cumming et al., 2013; Ostrom, 2007; Reid, Fernández-Giménez, & Galvin, 2014).

Our case study demonstrates that fundamental principles of SES theory and landscape governance theory are integral to but fading in traditional Maasai communities. The traditional grazing practices and land use practices, social networks and governance arrangements are changing but remain central in sustaining natural resource management from an ecosystem to landscape and regional level.

## 5.2 | Ecosystem scale

The pastoral lands of Kajiado and Narok counties in Kenya cover 40,000 km<sup>2</sup> along the border with Tanzania. The traditional common grazing grounds were divided into group ranches under the Group Ranch Representatives Act (CAP 287) in the 1960s and 1970s. Two of the group ranches in the southern Rift Valley of Kajiado County, Shompole and Olkiramatian, cover 1,500 km<sup>2</sup> (Figure 1) stretching from the forested Nguruman Escarpment to the semi-arid grasslands and bushland flats of the Rift Valley floor. The pre-colonial occupants of the Shompole–Olkiramatian ecosystem are predominantly members of the Lodokilani section of Maasai. The section includes a small group of settled farmers along the base of Ngurman Escarpment and a transhumant population of approximately 20,000 pastoralists who move seasonally.

Shompole and Olkiramatian communities have maintained their traditional collective herding practices to avoid the subdivision and fragmentation of their common grazing grounds arising from the subdivision and privatization of group ranches (Mwangi & Ostrom, 2009; Russell, Tyrrell, & Western, 2018). The two group ranches have, however, adapted traditional husbandry practices and livelihoods in response to the emerging opportunities in the market economy and to conform to national land, governance and development policies. The adaptations include diversifying the traditional livestock economy through commercial livestock production, farming, wildlife and tourism enterprises and pursuing new opportunities through education, social services and rural development (J. Kamanga, pers. commun.).

Both Shompole and Olkiramatian incorporate national legal requirements in group ranch governance statutes by appointing government chiefs alongside traditionally selected age set leaders responsible for overseeing the use of common-pool resources. Among other functions, the group ranch committees regulate the movement of livestock and seasonal pastures (Russell et al., 2018; Tyrrell et al., 2017; Western, 2019) and arbitrate herding arrangements within and between clans and with adjoining group ranches and Maasai sections (ibid.). The traditional late-season pastures

double up as wildlife conservancies attracting tourism enterprises, including lodges, cultural visits and nature walks (J. Kamanga, pers. comm.). The group ranch committees also oversee livestock markets and guide development activities through land planning and zoning to ensure complementarity in land and resource use.

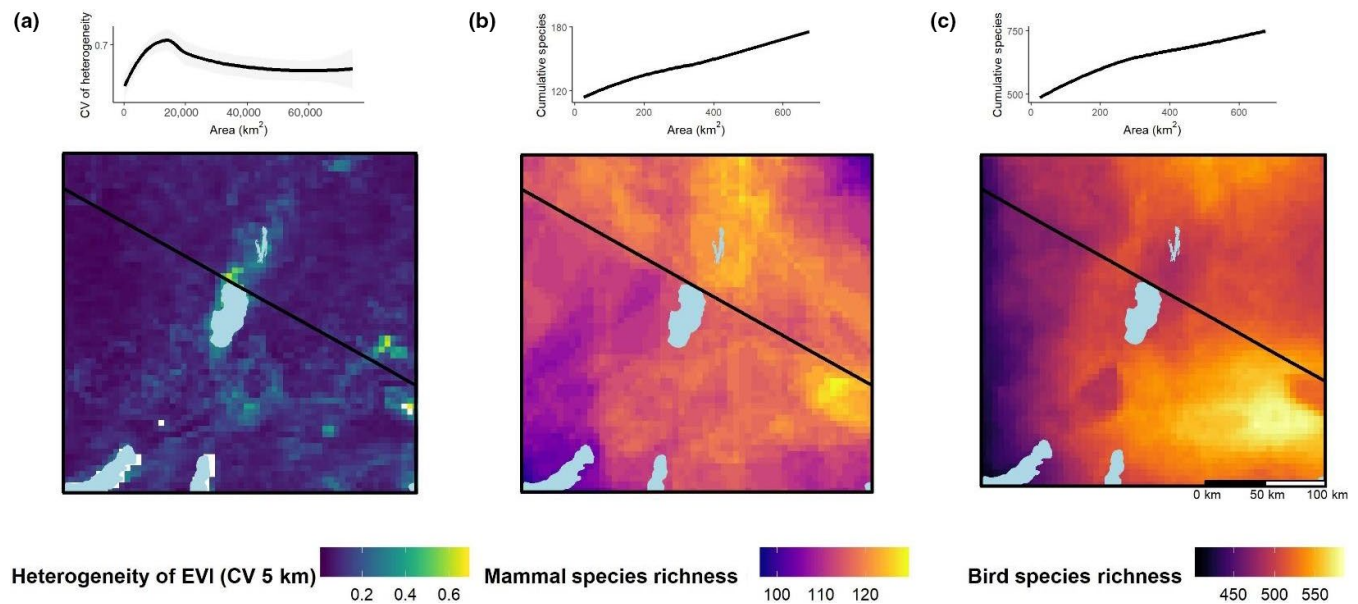
The regulation of livestock movements among community members ensures optimum use of pastures, the conservation of late-season grass reserves and minimum conflict with wildlife over pastures and water (Tyrrell et al., 2017). The governance practices afford community members access to grazing grounds, a variety of habitats for sustaining herd productivity and the mobility required to capitalize on patchy rains, avoid both disease outbreaks and competition with wildlife (Russell et al., 2018). The ability of herders to live alongside wildlife and benefit through tourism enterprises and traditional values (Roque de Pinho, 2009) depends on their rich knowledge of wild herbivores and carnivores and their skills in averting conflict (Western, 2019).

The role of traditional deployment of community scouts termed *ele'enore* is to gather the information the community needs to weigh its grazing options and reach collective decisions on herd deployment, watering regimes and the avoidance of serious conflicts with wildlife (Western, 2019). Olkiramatian, drawing on the *ele'enore* tradition, set up a Lale'enok Resource Centre which trains and deploys resources assessors to gather a wide variety of ecological, social and market data relevant to collective herd management, land planning, resource management and market access. The resource assessors use automated data collection platforms to enter and analyse information for rapid dissemination and decision-making (Mose & Western, 2015; Tyrrell et al., 2017).

The mix of traditional and contemporary knowledge, herding practices and governance arrangements maintains the seasonal scale of livestock movements on Shompole and Olkiramatian, ensuring that wildlife and livestock can benefit from seasonal migrations and adaptability to droughts (Russell et al., Tyrrell et al., 2017). In contrast to the sharp wildlife declines in pastoral areas which have been subdivided (Groom & Western, 2013; Ogutu et al., 2016; Said et al.), wildlife populations on Shompole and Olkiramatian have been more resilient. Elephant numbers have increased sharply, wild dog sightings have increased, cheetahs are resident and the lion population has spread into the surrounding ranches (Ahlering, Maldonado, Fleischer, Western, & Eggert, 2012; Schuette, Creel, & Christianson, 2013; Tyrrell et al., 2017). The area remains an important bird area with several threatened species (BirdLife International, 2019).<sup>2</sup>

## 5.3 | Landscape-scale

The success of the Olkiramatian and Shompole conservation efforts encouraged the African Conservation Centre (ACC) to create a landowner network connecting Amboseli National Park and Maasai Mara National Reserve across the Rift Valley, aimed at forging a new tourist link and conserving the biodiversity-rich landscape (Figures 1 and 3). ACC partnered with the Olkiramatian and Shompole leaders in 2007



**FIGURE 3** 5 × 5 km grid maps of the Kenya–Tanzania borderlands (border - black line) and major lakes (light blue) showing: (a) Habitat heterogeneity based on the coefficient of variation of Enhanced Vegetation Index (EVI) from Tuanmu and Jetz (2015). The accompanying graph shows heterogeneity increasing steeply with area coverage from a central pixel to a peak at 15,000 km<sup>2</sup>. The accompanying graphs to maps (b) and (c) show respectively species richness in mammals using the Digital Distribution Maps of Red List Species (IUCN, 2016) and birds based on Bird Life Digital Distribution Maps (BirdLife International, 2019) both increasing steadily from a central pixel

to establish the South Rift Association of Landowners (SORALO) incorporating the 16 current and former group ranches spanning the 15,000 km<sup>2</sup> landscape. The common goals of the SORALO members include land security, raising and diversifying livelihoods, sustaining natural resource use and land health, developing ecotourism enterprises and conserving the open rangelands and rich wildlife assemblage.

The governance of SORALO's membership builds on the existing social networks and reciprocal grazing arrangements connecting adjacent Maasai clans and sections. SORALO is registered as a land trust and like Shompole and Olkiramatian, incorporates traditional and contemporary governance institutions and practices. The SORALO board includes representatives of regional group ranch clusters and draws on strong community inputs and the political influence it has with county and national agencies.

The geographic reach and group ranch network SORALO covers scales up Shompole and Olkiramatian ecosystem-level governance to take advantage of the expanded benefits accruing from a landscape scale, including access to drought refuges, the conservation of watersheds and ecosystem services they depend on beyond their boundaries, the protection of wildlife corridors, joint wildlife scouting operations, ecotourism planning and access to country, national government and NGO services.

## 5.4 | Regional-scale

The Kenya–Tanzania borderlands are globally important as a vertebrate biodiversity hotspot and regionally as a centre of plant

diversity accounting for a quarter of all the plants recorded in Kenya and Tanzania (Ministry of Environment Natural Resources and Regional Development Authorities, 2015; Figure 3). The borderlands also support large populations of endangered and threatened species, including one of the largest free-ranging elephant populations, richest carnivore assemblages in Africa, and among the largest remaining large herbivore migrations world-wide (Harris, Thirgood, Hopcraft, Croomsigt, & Berger, 2009; Jenkins et al., 2013). Coupled with sixteen protected areas and a \$1.5 billion annual tourism industry, the borderlands have attracted international attention as one of the earth's last great natural wonders.

Despite the attention and large well-equipped government ranger forces, wildlife has declined sharply in national parks both sides of the border in the last few decades (Craigie et al., 2010; Western, Russell, & Cuthill, 2009). The declines stem from the same threats confronting the future of free-ranging pastoralists at an ecosystem and group ranch level; land conversion, subdivision, settlement and growing pressures on the open rangelands. Species dependent on large-scale migrations to access seasonal forage and sustain large populations are especially vulnerable to land compression and fragmentation. Both the Tarangire and Nairobi National Park herbivore populations are threatened by a lack of collaboration and incentives at the regional scale (Newmark, 2008; Western & Gichohi, 1993).

In the case of elephants, the main cause of the decline in the 1970s and 1980s arose from a ten-fold rise in ivory prices in the world market creating a surge in poaching and compression of remaining herds into the relative safety of national parks (Mose & Western, 2015; Stiles, 2004). The compression reduced woody



habitats and biodiversity in national parks and most severely in small parks like Amboseli. Here, the extreme compression and habitat loss (Western & Maitumo, 2004) spurred efforts in 2006 to reestablish the elephant's range to reduce the pressure on the park and recolonize the elephant's former wide-ranging movements. The first step involved training and deploying community scouts and promoting conservancies and tourism enterprises in Shompole and Olkiramatian where resident herds were driven out by poachers in the 1980s.

Following the natural repopulation of Shompole and Olkiramatian, ACC and SORALO expanded the elephant recovery program to connect the isolated Mara and Amboseli population with the aim of creating a metapopulation across the Rift Valley. The recovery involved monitoring elephant movements and deploying community scouts along dispersal routes. The program, premised on the inside-out model of conservation, evolved into the BCI in 2012.

The initial goal of BCI was to halt elephant poaching by using SORALO-like community linkages to create space for the fragmented park populations and recreate a viable metapopulation in the pastoral lands spanning the Kenya-Tanzania borderlands. BCI used the endangerment and charismatic appeal of elephants to create a coalition of local NGOs, conservation organizations, researchers and government agencies to support the pastoral community initiatives. The BCI collaboration soon expanded from the initial focus on elephants to conserving lions, other large carnivores, giraffe and other threatened and charismatic species to obviate the impacts of land restrictions on large scale movements and the fragmentation of metapopulations critical to their genetic and ecological viability. Recent evidence from the Serengeti-Mara ecosystem shows that even the largest of national parks are threatened by surrounding human pressures (Veldhuis et al., 2019) and are in need of mitigating conservation measures aimed at sustaining free-ranging movements.

## 5.5 | Inside-out conservation as a way forward

Conservation in working landscapes connected from small land management entities to large multiscale landscapes holds untapped potential for alleviating the insularization and other human impacts on protected areas as well as conserving biodiversity in rural landscapes. Global warming is adding urgency to conserving species beyond protected areas as their ranges shift in response to climate change (Thomas et al., 2004). The threats call for new approaches to conserving biodiversity in the rural landscape. Solutions range from creating buffer zones for protected areas, protected areas that include working landscapes, conservation easements, leases and direct land purchases. Conservation opportunities in the agrarian landscape include creating space through land-sparing intensified agriculture; conservation subsidies to farmers to protect wetlands, rivers and woodlands and financial offsets for converting inimical uses to conservation-compatible uses.

Conservation emanating from within communities based on sustaining livelihoods which are dependent on large open landscapes widens the scope for nature conservation in the rural lands which cover 75% of the earth's terrestrial surface. Contrary to both top-down and bottom-up approaches which are focused primarily on conserving wildlife as the priority, an inside-out approach builds on existing practices that maintain open landscapes and wildlife by patching together networks of livelihoods and interests with compatible objectives. Unlike debates around land sparing and land sharing, this approach is centered on people rather than wildlife (Phalan, 2018) and allows for a range of land use options along the sharing to sparing continuum. As our case study shows, the productivity and resilience of pastoral communities are linked to the ecological and functional benefits of scale, heterogeneity and land health, which if sustained, make space for large free-ranging wildlife populations and biodiversity. While the inside-out approach is applicable to rangelands, it has relevance to other large complex systems that benefit from increasing the scale of management to larger landscapes, such as marine fisheries (Curtin, 2015).

Finding such space also depends on finding a place for biodiversity within a community's aspirations and economic options. To succeed, biodiversity conservation in rural lands must expand to a landscape scale and cut across national, jurisdictional, institutional and cultural boundaries, and create network connections between them (Scarlett & McKinney, 2016). The feasibility of using an inside-out approach to conserve large open spaces becomes more feasible when linked to existing theories connecting landscapes among multiple jurisdictions using a polycentric and devolved approach to governance (Ostrom, 2007), landscape theory based on sustainable practices (Arts et al., 2017; Curtin, 2015) and a sound ecological understanding using functional heterogeneity theory (Fuhlendorf, Fynn, McGranahan, & Twidwell, 2017; Fynn et al., 2016).

Of particular use is the SES framework (McGinnis & Ostrom, 2014; Ostrom, 2007, 2009), which allows for the analysis of relationships among multiple levels of complex SESs at different spatial and temporal scales. It also helps us understand how specific parts of the systems are related and interact. Applying the SES framework to natural resource management reveals several important elements affect the likelihood of users' self-organizing to sustainably manage a resource, including: (a) the existence of institutions that work at the correct social and ecological scales; (b) governance that is multi-scale and multi-level, (c) communities who have clear devolved autonomy over resource management; (d) communities who see an importance of the resource, and see a benefit (through economic and non-economic values) from their natural resources; (e) strong social norms of collaborative governance and management (Cumming et al., 2013; Ostrom, 2007, 2009; Reid et al., 2014).

Our case study shows that these conditions are achievable for the management of pasture (and subsequently wildlife), and when they are conceptually twinned with a capacity for effective landscape

governance (see Arts et al., 2017 for landscape governance as a capacity; Figure 1) and locally appropriate ecological principles, the results are mutual benefits to people, livestock and wildlife.

Biodiversity and livelihoods dependent on land health converge in the rangelands more than in most other biomes. The convergence stems from two formerly disparate objectives. First, the ambit of protected areas for conserving biodiversity has widened from the strict protectionism of IUCN Category I parks to include Category VI areas accommodating a range of compatible human values and uses. Second, land users, including pastoralists, private commercial ranchers, fishing, hunting and forestry communities are finding common ground in conserving healthy, open landscape for their inclusive values. These include not only primary livelihoods but also other uses and values as nature sensibilities expand and international commitments to sustainable development deepen. These include a sense of place, a healthy environment and the range of ecological services it provides (Arts et al., 2017; Sayer et al., 2013).

The inside-out approach to winning space and a place for wildlife and biodiversity is place-based and draws on local knowledge and informal governance arrangements. It avoids the bottom-up wildlife-centric approaches which are driven from the outside and that alienate communities working on the land for a living. Instead, conservation approached from the inside outward promotes a human-centered approach that reinforces land health and spatial connectivity. It also encourages a shift from colonial and post-colonial central government command-and-control policies that have hampered wildlife and natural resource management towards devolved rights and ownership emerging with political and economic liberalization trends around the world.

Expanding conservation benefits from the inside-out is especially pertinent to conserving large herbivores and carnivores which need large open landscapes, pose considerable threats and costs to rural communities and are the most vulnerable of all species to farming, ranching and resource extraction (Tilman et al., 2017). Given that the downsizing of large herbivores and carnivores will have the biggest impact on ecosystem structure and function in coming decades, finding places for such landscape species will avert trophic cascade effects and the ecological impoverishment arising from the loss of keystone species (Dirzo et al., 2014; Smith et al., 2018).

Unlike much of East Africa's privately protected areas, conservancies and Wildlife Management Areas which 'buy' tolerance through tourism, trophy hunting or other land leases, the inside-out approach to conserving large open landscapes and wildlife expands self-interests and shared community livelihoods, social networks and cultural values. In the case of the Maasai, *eramatare* linkages at the individual levels between family and herds, and at the community level for land access and land health through common pool governance arrangements, these connections promote both traditional intrinsic and instrumental values for wildlife. The linkages also reduce conflict and increase the prospects for coexistence with wildlife. Nevertheless, governments, NGOs and other conservation agencies have a large supporting role to play in

offsetting the opportunity costs of wildlife and mitigating conflict as subsistence economies erode with population growth and market economies.

Collaboration at large scales and across human-dominated landscapes faces enormous challenges and at best will only complement and not supplant the need for protected areas and other conservation tools. Yet, as Leopold recognized, the extension of ethical values to include the health of the land is an evolutionary possibility and ecological necessity. By drawing on lessons from cultural institutions, principles and practices which have maintained the commons for generations, combined with ecological, political and social sciences, we see the possibility of landscape doctoring evolving into the science of land health.

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## CONFLICT OF INTEREST

J.K., G.W. and S.R. work for SOLARO and D.W. works for the African Conservation Centre. All authors have engaged in conservation within the region outlined in the case study.

## AUTHORS' CONTRIBUTIONS

D.W., S.R. and P.T. developed the original concept for this manuscript. D.W. led the writing. P.T. and P.B. produced the figures and analysed the data and assisted D.W. with the drafting of the manuscript alongside S.R., G.W. and J.K. All authors revised the manuscript critically for important intellectual content, approved the version to be published and agree to be accountable for the aspects of the work that they conducted.

## DATA AVAILABILITY STATEMENT

All data in this manuscript are available through their referenced sources.

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## ENDNOTES

<sup>1</sup> [https://www.accafrica.org/our\\_work/explore\\_programs/conserving-biodiversity-in-east-africa/kenya-tanzania-borderland-conservation-initiative/](https://www.accafrica.org/our_work/explore_programs/conserving-biodiversity-in-east-africa/kenya-tanzania-borderland-conservation-initiative/)

<sup>2</sup> BirdLife International (2019) Important Bird Areas factsheet: South Nguruman. Retrieved from <http://www.birdlife.org> and [http://www.conservationleadershipprogramme.org/media/2014/11/000703\\_Kenya\\_FR\\_BirdsSouthNguruman.pdf](http://www.conservationleadershipprogramme.org/media/2014/11/000703_Kenya_FR_BirdsSouthNguruman.pdf)

## REFERENCES

- Adams, M. (2005). Beyond Yellowstone? Conservation and Indigenous rights in Australia and Sweden. In G. Cant, A. Goodall, & J. Inns (Eds.), *Discourses and silences: Indigenous peoples, risks and resistance* (pp. 127–138). Christchurch, New Zealand: Department of Geography, University of Canterbury.
- Adams, J. S., & McShane, T. O. (1992). *The myth of wild Africa: Conservation without illusion*. Berkeley, CA: University of California Press.
- Adler, R. W. (2007). *Restoring Colorado river ecosystems*. Washington, DC: Island Press.
- Ahlering, M. A., Maldonado, J. E., Fleischer, R. C., Western, D., & Eggert, L. S. (2012). Fine-scale group structure and demography of African savanna elephants recolonizing lands outside protected areas. *Diversity and Distributions*, 18(10), 952–961. <https://doi.org/10.1111/j.1472-4642.2012.00896.x>
- Arts, B., Buizer, M., Horlings, L., Ingram, V., Van, O. C., & Opdam, P. (2017). Landscape approaches: A state-of-the-art review. *Annual Review of Environment and Resources*, 42, 439–463.
- Ash, A., Gross, J., & Smith, M. S. (2004). Scale, heterogeneity and secondary production in tropical rangelands. *African Journal of Range & Forage Science*, 21, 137–145. <https://doi.org/10.2989/10220110409485846>
- Asner, G. P., Vaughn, N., Smit, I. P. J., & Levick, S. (2016). Ecosystem-scale effects of megafauna in African savannas. *Ecography*, 39, 240–252. <https://doi.org/10.1111/ecog.01640>
- Assessment M. E. (2003). *Millennium Ecosystem Assessment (MA): Strengthening capacity to manage ecosystems sustainably for human well-being*. Washington, DC: World Resources Institute.
- Bedelian, C., & Ogotu, J. O. (2017). Trade-offs for climate-resilient pastoral livelihoods in wildlife conservancies in the Mara ecosystem, Kenya. *Pastoralism*, 7, 1–22. <https://doi.org/10.1186/s13570-017-0085-1>
- Bennett, A. F., Haslem, A., Cheal, D. C., Clarke, M. F., Jones, R. N., Koehn, J. D., ... Yen, A. L. (2009). Ecological processes: A key element in strategies for nature conservation. *Ecological Management and Restoration*, 10, 192–199. <https://doi.org/10.1111/j.1442-8903.2009.00489.x>
- BirdLife International. (2019). BirdLife Data Zone. Retrieved from [www.datazone.birdlife.org](http://www.datazone.birdlife.org)
- Boone, R. B. (2005). Quantifying changes in vegetation in shrinking grazing areas in Africa. *Conservation and Society*, 3, 150–173.
- Brockington, D., Adams, W. M., Agrawal, B., Agrawal, A., Büscher, B., Chatre, A., ... Oldekop, J. A. (2019). Working governance for working land. *Science*, 362, 1257.
- Büscher, B. F. R., & Brockington, D. (2017). Half earth or whole earth? Radical ideas for conservation, and their implications. *Oryx*, 12, 145.
- Callicott, J. B. (1990). Whither conservation ethics? *Conservation Biology*, 4, 15–20. <https://doi.org/10.1111/j.1523-1739.1990.tb00261.x>
- Costanza, R., d'Arge, R., de Groot, R., Farber, S., Grasso, M., Hannon, B., ... van den Belt, M. (1997). The value of the world's ecosystem services and natural capital. *Nature*, 387, 253–260. <https://doi.org/10.1038/387253a0>
- Craigie, I. D., Baillie, J. E. M., Balmford, A., Carbone, C., Collen, B., Green, R. E., & Hutton, J. M. (2010). Large mammal population declines in Africa's protected areas. *Biological Conservation*, 143, 2221–2228. <https://doi.org/10.1016/j.biocon.2010.06.007>
- Cumming, D. H. M. (2008). *Large scale conservation planning and priorities for the Kavango-Zambezi Transfrontier Conservation Area*. Unpublished report commissioned by Conservation International.
- Cumming, G. S., Olsson, P., Chapin, F. S., & Holling, C. S. (2013). Resilience, experimentation, and scale mismatches in social-ecological landscapes. *Landscape Ecology*, 28, 1139–1150. <https://doi.org/10.1007/s10980-012-9725-4>
- Curtin, C. G. (2002). Integration of science and community-based conservation in the Mexico/US borderlands. *Conservation Biology*, 16, 880–886. <https://doi.org/10.1046/j.1523-1739.2002.00165.x>
- Curtin, C. G. (2015). *The science of open spaces: Theory and practice for conserving large, complex systems*. Washington, DC: Island Press.
- de Pinho, J. R., & Ellis, J. (2009). *Staying together: People-wildlife relationship in the Amboseli Ecosystem, southern Kenya*. Davis, CA: GL-CRSP.
- Dinerstein, E., Olson, D., Joshi, A., Vynne, C., Burgess, N. D., Wikramanayake, E., ... Saleem, M. (2017). An ecoregion-based approach to protecting half the terrestrial realm. *BioScience*, 67, 534–545. <https://doi.org/10.1093/biosci/bix014>
- Dirzo, R., Young, H. S., Galetti, M., Ceballos, G., Isaac, N. J. B., & Collen, B. (2014). Defaunation in the anthropocene. *Science*, 401, 401–406.
- Dolrenry, S., Stenglein, J., Hazzah, L., Lutz, R. S., & Frank, L. (2014). A metapopulation approach to African lion (*Panthera leo*) conservation. *PLoS ONE*, 9, e88081.
- Donaldson, L., Wilson, R. J., & Maclean, I. M. D. (2017). Old concepts, new challenges: Adapting landscape-scale conservation to the twenty-first century. *Biodiversity and Conservation*, 26, 527–552. <https://doi.org/10.1007/s10531-016-1257-9>
- Dudley, N., & Stolton, S. (2008). *Defining protected areas: An international conference in Almeria, Spain*. Gland, Switzerland: IUCN.
- Fischer, J., Abson, D. J., Butsic, V., Chappell, M. J., Ekroos, J., Hanspach, J., ... von Wehrden, H. (2014). Land sparing versus land sharing: Moving forward. *Conservation Letters*, 7, 149–157. <https://doi.org/10.1111/conl.12084>
- Fryxell, J. M., Wilmschurst, J. F., Sinclair, A. R. E., Haydon, D. T., Holt, R. D., & Abrams, P. A. (2005). Landscape scale, heterogeneity, and the viability of Serengeti grazers. *Ecology Letters*, 8, 328–335. <https://doi.org/10.1111/j.1461-0248.2005.00727.x>
- Fuhlendorf, S. D., Fynn, R. W. S., McGranahan, D. A., & Twidwell, D. (2017). Heterogeneity as the basis for rangeland management. In *Rangeland systems* (pp. 169–196). New York, NY: Springer.
- Fuhlendorf, S. D., Harrell, W. C., Engle, D. M., Hamilton, R. G., Davis, C. A., & Leslie Jr, D. M. (2016). Should heterogeneity be the basis for conservation? Grassland bird response to fire and grazing. *Ecological Applications*, 16, 1706–1716. [https://doi.org/10.1890/1051-0761\(2006\)016\[1706:shbtbf\]2.0.co;2](https://doi.org/10.1890/1051-0761(2006)016[1706:shbtbf]2.0.co;2)
- Fynn, R. W. S., Augustine, D. J., Peel, M. J. S., & de Garine-Wichatitsky, M. (2016). Review: Strategic management of livestock to improve biodiversity conservation in African savannahs: A conceptual basis for wildlife-livestock coexistence. *Journal of Applied Ecology*, 53, 388–397.
- Fynn, R. W. S., & Bonyongo, M. C. (2011). Functional conservation areas and the future of Africa's wildlife. *African Journal of Ecology*, 49(2), 175–188. <https://doi.org/10.1111/j.1365-2028.2010.01245.x>
- Fynn, R. W. S., Murray-Hudson, M., Dhliwayo, M., & Scholte, P. (2015). African wetlands and their seasonal use by wild and domestic herbivores. *Wetlands Ecology and Management*, 23, 559–581. <https://doi.org/10.1007/s11273-015-9430-6>
- Groom, R., & Western, D. (2013). Impact of land subdivision and sedentarization on wildlife in Kenya's southern rangelands. *Rangeland Ecology & Management*, 66(1), 1–9. <https://doi.org/10.2307/23355261>
- Hamilton, M., Milne, B., Walker, R., & Brown, J. H. (2007). Nonlinear scaling of space use in human hunter-gatherers. *Philosophical Transactions of the Royal Society B: Biological Sciences*. <https://doi.org/10.1073/pnas.0611197104>
- Harris, G., Thirgood, S., Hopcraft, J. G. C., Cromsigt, J. P. G. M., & Berger, J. (2009). Global decline in aggregated migrations of large terrestrial mammals. *Endangered Species Research*, 7, 55–76. <https://doi.org/10.3354/esr00173>
- Hobbs, N. T., Galvin, K. A., Stokes, C. J., Lockett, J. M., Ash, A. J., Boone, R. B., ... Thornton, P. K. (2008). Fragmentation of rangelands: Implications for humans, animals, and landscapes. *Global Environmental Change*, 18, 776–785. <https://doi.org/10.1016/j.gloenvcha.2008.07.011>
- Illius, A. W., & O'Connor, T. G. (2000). Resource heterogeneity and ungulate population dynamics. *Oikos*, 89, 283–294.
- IUCN. (2016). The IUCN Red List of Threatened Species. Version 2016-3. Retrieved from <http://www.iucnredlist.org>
- Jasch, C. (2003). The use of Environmental Management Accounting (EMA) for identifying environmental costs. *Journal of Cleaner*

- Production*, 11, 667–676. [https://doi.org/10.1016/S0959-6526\(02\)00107-5](https://doi.org/10.1016/S0959-6526(02)00107-5)
- Jenkins, C. N., Pimm, S. L., & Joppa, L. N. (2013). Global patterns of terrestrial vertebrate diversity and conservation. *Proceedings of the National Academy of Sciences of the United States of America*, 110(28), E2602–E2610. <https://doi.org/10.1073/pnas.1302251110>
- Jones, K. R., Venter, O., Fuller, R. A., Allan, J. R., Maxwell, S. L., Negret, P. J., & Watson, J. E. M. (2018). One-third of global protected land is under intense human pressure. *Science*, 360, 788–791. <https://doi.org/10.1126/science.aap9565>
- Joppa, L. N., & Pfaff, A. (2009). High and far: Biases in the location of protected areas. *PLoS ONE*, 4, 1–6.
- Kremen, C., & Merenlender, A. M. (2018). Landscapes that work for biodiversity and people. *Science*, 362, eaau6020. <https://doi.org/10.1126/science.aau6020>
- Kristjanson, P., & Trench, P. C. (2009). Changing land use, livelihoods and wildlife conservation in Maasailand. In K. Homewood, P. Kristjanson, & P. Trench (Eds.), *Staying maasai?* (pp. 1–42). New York, NY: Springer.
- KWCA. (2016). *State of wildlife conservancies in Kenya report 2016* (pp. 1–84). Nairobi, Kenya: KWCA.
- Leopold, A. (1949). *A sand county Almanac*. New York, NY: Ballantine.
- McGinnis, M. D., & Ostrom, E. (2014). Social-ecological system framework: Initial changes and continuing challenges. *Ecology and Society*, 19, <https://doi.org/10.5751/es-06387-190230>
- McNeely, J. A. (1993). Economic incentives for conserving biodiversity: Lessons for Africa. *Ambio*, 22, 144–150.
- Meine, C., Soulé, M. E., & Noss, R. (2006). A mission-driven discipline. *Conservation Biology*, 20(631), 651.
- Merrill, T. (2005). Conservation Strategy for Grizzly Bears in the Yellowstone to Yukon Ecoregion. Yellowstone to Yukon Conservation Initiative (Vol. 6). Technical Report.
- Ministry of Environment Natural Resources and Regional Development Authorities. (2015). *Kenyas natural capital A biodiversity Atlas*. Nairobi, Kenya: Ministry of Environment Natural Resources and Regional Development Authorities.
- Morrison, T. A., Archibald, S., Michael, T., Dobson, A. P., Donaldson, J. E., Gareth, P., & Probert, J. (2018). Pyrodiversity interacts with rainfall to increase bird and mammal richness in African savannas. *Ecology Letters*, 21, 557–567.
- Mose, V. N., Nguyen-Huu, T., Western, D., Auger, P., & Nyandwi, C. (2013). Modelling the dynamics of migrations for large herbivore populations in the Amboseli National Park, Kenya. *Ecological Modelling*, 254, 43–49. <https://doi.org/10.1016/j.ecolmodel.2013.01.016>
- Mose, V. N., & Western, D. (2015). Spatial cluster analysis for large herbivore distributions: Amboseli ecosystem, Kenya. *Ecological Informatics*, 30, 203–206. <https://doi.org/10.1016/j.ecoinf.2015.05.010>
- Muchiru, A. N., Western, D., & Reid, R. S. (2009). The impact of abandoned pastoral settlements on plant and nutrient succession in an African savanna ecosystem. *Journal of Arid Environments*, 73, 322–331. <https://doi.org/10.1016/j.jaridenv.2008.09.018>
- Mwangi, E., & Ostrom, E. (2009). Top-down solutions: Looking up from east Africa's rangelands. *Environment*, 9157, 37–41.
- Nabhan, G. P. (2012). *Where our food comes from: Retracing Nikolay Vavilov's quest to end famine*. Washington, DC: Island Press.
- Naidoo, R., Weaver, L. C., Diggle, R. W., Matongo, G., Stuart-Hill, G., & Thouless, C. (2016). Complementary benefits of tourism and hunting to communal conservancies in Namibia. *Conservation Biology*, 30, 628–638. <https://doi.org/10.1111/cobi.12643>
- Nash, R. F. (1989). *The rights of nature: A history of environmental ethics*. Madison, WI: University of Wisconsin Press.
- Nelson, F., Foley, C., Foley, L. S., Leposo, A., Loure, E., Peterson, D., ... Williams, A. (2010). Payments for ecosystem services as a framework for community based conservation in northern Tanzania. *Conservation Biology*, 24, 78–85. <https://doi.org/10.1111/j.1523-1739.2009.01393.x>
- Newbold, T. (2018). Future effects of climate and land-use change on terrestrial vertebrate community diversity under different scenarios. *Proceedings of the Royal Society B: Biological Sciences*, 285. <https://doi.org/10.1098/rspb.2018.0792>
- Newmark, W. D. (2008). Isolation of African protected areas. *Frontiers in Ecology and the Environment*, 6, 321–328.
- Ogutu, J. O., Piepho, H.-P., Said, M. Y., Ojwang, G. O., Njino, L. W., Kifugo, S. C., & Wargute, P. W. (2016). Extreme wildlife declines and concurrent increase in livestock numbers in Kenya: What are the causes? *PLoS ONE*, 11, e0163249. <https://doi.org/10.1371/journal.pone.0163249>
- Oldekop, J. A., Holmes, G., Harris, W. E., & Evans, K. L. (2016). A global assessment of the social and conservation outcomes of protected areas. *Conservation Biology*, 30, 133–141. <https://doi.org/10.1111/cobi.12568>
- Osipova, L., Okello, M. M., Njumbi, S. J., Ngene, S., Western, D., Hayward, M. W., & Balkenhol, N. (2018). Using step-selection functions to model landscape connectivity for African elephants: Accounting for variability across individuals and seasons. *Animal Conservation*, 22(1), 35–48. <https://doi.org/10.1111/acv.12432>
- Ostrom, E. (2007). A diagnostic approach for going beyond panaceas. *Proceedings of the National Academy of Sciences of the United States of America*, 104, 15181–15187. <https://doi.org/10.1073/pnas.0702288104>
- Ostrom, E. (2009). A general framework for analyzing sustainability of social-ecological systems. *Science*, 325, 419–422. <https://doi.org/10.1126/science.1172133>. Retrieved from <http://science.sciencemag.org/content/325/5939/419>
- Owen-Smith, N. (2004). Functional heterogeneity in resources within landscapes and herbivore population dynamics. *Landscape Ecology*, 19, 761–771. <https://doi.org/10.1007/s10980-005-0247-2>
- Owen-Smith, R. N. (2002). *Adaptive herbivore ecology: From resources to populations in variable environments*. Cambridge, UK: Cambridge University Press.
- Peterson, G., Allen, C., & Holling, C. S. (1998). Ecological resilience, biodiversity, and scale. *Ecosystems*, 1, 6–18.
- Phalan, B. T. (2018). What have we learned from the land sparing-sharing model? *Sustainability (Switzerland)*, 10, 1–24. <https://doi.org/10.3390/su10061760>
- Pressey, R. L., & Bottrill, M. C. (2008). Opportunism, threats, and the evolution of systematic conservation planning. *Conservation Biology*, 22, 1340–1345. <https://doi.org/10.1111/j.1523-1739.2008.01032.x>
- Pressey, R. L., Visconti, P., & Ferraro, P. J. (2015). Making parks make a difference: Poor alignment of policy, planning and management with protected-area impact, and ways forward. *Philosophical Transactions of the Royal Society B: Biological Sciences*, 370, 20140280.
- Reid, R. S., Fernández-Giménez, M. E., & Galvin, K. A. (2014). Dynamics and resilience of rangelands and pastoral peoples around the globe. *Annual Review of Environment and Resources*, 39, 217–242. <https://doi.org/10.1146/annurev-environ-020713-163329>
- Roque de Pinho, J. (2009). *Staying together: People-wildlife relationships in a pastoralist society in transition, Amboseli Ecosystem, southern Kenya*. Fort Collins, CO: Colorado State University.
- Russell, S., Tyrrell, P., & Western, D. (2018). Seasonal interactions of pastoralists and wildlife in relation to pasture in an African savanna ecosystem. *Journal of Arid Environments*, 154, 70–81. <https://doi.org/10.1016/j.jaridenv.2018.03.007>
- Said, M. Y., Ogutu, J. O., Kifugo, S. C., Makui, O., Reid, R. S., & de Leeuw, J. (2016). Effects of extreme land fragmentation on wildlife and livestock population abundance and distribution. *Journal for Nature Conservation*, 34, 151–164. <https://doi.org/10.1016/j.jnc.2016.10.005>
- Sayer, J., Sunderland, T., Ghazoul, J., Pfund, J.-L., Sheil, D., Meijaard, E., ... Buck, L. E. (2013). Ten principles for a landscape approach to reconciling agriculture, conservation, and other competing land uses. *Proceedings of the National Academy of Sciences of the United States of America*, 110, 8349–8356. <https://doi.org/10.1073/pnas.1210595110>

- Scarlett, L., & McKinney, M. (2016). Connecting people and places: The emerging role of network governance in large landscape conservation. *Frontiers in Ecology and the Environment*, 14, 116–125. <https://doi.org/10.1002/fee.1247>
- Schuette, P., Creel, S., & Christianson, D. (2013). Coexistence of African lions, livestock, and people in a landscape with variable human land use and seasonal movements. *Biological Conservation*, 157, 148–154. <https://doi.org/10.1016/j.biocon.2012.09.011>
- Shahabuddin, G., & Rao, M. (2010). Do community-conserved areas effectively conserve biological diversity? Global insights and the Indian context. *Biological Conservation*, 143, 2926–2936. <https://doi.org/10.1016/j.biocon.2010.04.040>
- Sitarz, D. (1993). *Agenda 21: The earth summit strategy to save our planet*. Boulder, CO: EarthPress.
- Smith, F. A., Smith, R. E. E., Lyons, S. K., & Payne, J. L. (2018). Body size downgrading of mammals over the late Quaternary. *Science*, 313, 310–313. <https://doi.org/10.1126/science.aao5987>
- Spear, T., & Waller, R. (1993). *Being Maasai*. London, UK: James Currey.
- Stiles, D. (2004). The ivory trade and elephant conservation. *Environmental Conservation*, 31, 309–321. <https://doi.org/10.1017/S0376892904001614>
- Thomas, K. (1983). *Man and the natural world: Changing attitudes in England, 1500–1800*. London, UK: Allen Lane.
- Thomas, C. D., Cameron, A., Green, R. E., Bakkenes, M., Beaumont, L. J., Collingham, Y. C., & Erasmus, B. F. N. (2004). Extinction risk from climate change. *Nature*, 427, 145–148.
- Tilman, D., Clark, M., Williams, D. R., Kimmel, K., Polasky, S., & Packer, C. (2017). Future threats to biodiversity and pathways to their prevention. *Nature*, 546, 73. <https://doi.org/10.1038/nature22900>
- Tuanmu, M.-N., & Jetz, W. (2015). A global, remote sensing-based characterization of terrestrial habitat heterogeneity for biodiversity and ecosystem modelling. *Global Ecology and Biogeography*, 24(11), 1329–1339. <https://doi.org/10.1111/geb.12365>
- Tyrrell, P., Russell, S., & Western, D. (2017). Seasonal movements of wildlife and livestock in a heterogeneous pastoral landscape: Implications for coexistence and community based conservation. *Global Ecology and Conservation*, 12, 59–72. <https://doi.org/10.1016/j.gecco.2017.08.006>
- United Nations. (1992). *Convention on biological diversity*. Diversity:30. Retrieved from <http://www.cbd.int/doc/legal/cbd-en.pdf>
- Valls-Fox, H., Chamailé-Jammes, S., de Garine-Wichatitsky, M., Perrotton, A., Courbin, N., Miguel, E., ..., Muzamba, M. (2018). Water and cattle shape habitat selection by wild herbivores at the edge of a protected area. *Animal Conservation*, 21, 365–375.
- Veldhuis, M. P., Ritchie, M. E., Ogotu, J. O., Morrison, T. A., Beale, C. M., Estes, A. B., ... Ojwang, G. O. (2019). Cross-boundary human impacts compromise the Serengeti-Mara ecosystem. *Science*, 363, 1424–1428.
- Venter, O., Magrath, A., Outram, N., Klein, C. J., Possingham, H. P., Di Marco, M., & Watson, J. E. M. (2017). Bias in protected-area location and its effects on long-term aspirations of biodiversity conventions. *Conservation Biology*, 32, 127–134. <https://doi.org/10.1111/cobi.12970>
- Vucetich, J. A., Bruskotter, J. T., & Nelson, M. P. (2015). Evaluating whether nature's intrinsic value is an axiom or anathema to conservation. *Conservation Biology*, 29, 321–332. <https://doi.org/10.1111/cobi.12464>
- Vuorio, V., Muchiru, A., Reid, R. S., & Ogotu, J. O. (2014). How pastoralism changes savanna vegetation impact of old pastoral settlements on plant diversity and abundance in south western Kenya. *Biodiversity and Conservation*, 23, 3219–3240. <https://doi.org/10.1007/s10531-014-0777-4>
- Wang, G., Hobbs, N. T., Boone, R. B., Illius, A. W., Gordon, I. J., Gross, J. E., & Hamlin, K. L. (2006). Spatial and temporal variability modify density dependence in populations of large herbivores. *Ecology*, 87, 95–102. <https://doi.org/10.1890/05-0355>
- Watson, J. E. M., Dudley, N., Segan, D. B., & Hockings, M. (2014). The performance and potential of protected areas. *Nature*, 515, 67–73. <https://doi.org/10.1038/nature13947>. Retrieved from <http://www.nature.com/nature/journal/v515/n7525/pdf/nature13947.pdf>
- West, G. (2017). *Scale: The universal laws of growth, innovation, sustainability, and the pace of life in organisms, cities, economies, and companies*. New York, NY: Penguin Press.
- Western, D. (1982). The environment and ecology of pastoralists in Arid Savannas. *Development and Change*, 13, 183–211. <https://doi.org/10.1111/j.1467-7660.1982.tb00117.x>
- Western, D. (1997). *In the dust of Kilimanjaro*. Washington, DC: Island Press.
- Western, D., & Finch, V. (1986). Cattle and pastoralism: Survival and production in arid lands. *Human Ecology*, 14, 77–94. <https://doi.org/10.1007/BF00889211>
- Western, D., & Gichohi, H. (1993). Segregation effects and the impoverishment of savanna parks: The case for ecosystem viability analysis. *African Journal of Ecology*, 31, 269–281. <https://doi.org/10.1111/j.1365-2028.1993.tb00541.x>
- Western, D., & Maitumo, D. (2004). Woodland loss and restoration in a savanna park. *African Journal of Ecology*, 42, 111–121.
- Western, D., Mose, V., Worden, J., & Maitumo, D. (2015). Predicting extreme droughts in savannah africa: A comparison of proxy and direct measures in detecting biomass fluctuations, trends and their causes. *PLoS ONE*, 10, e0136516. <https://doi.org/10.1371/journal.pone.0136516>
- Western, D., & Nightingale, M. (2005). Keeping the East African rangelands open and productive. *Conservation and People*, 1, 1–8.
- Western, D., & Pearl, M. (1989). *Conservation for the 21st century*. New York, NY: Oxford University Press.
- Western, D., Russell, S., & Cuthill, I. (2009). The status of wildlife in protected areas compared to non-protected areas of Kenya. *PLoS ONE*, 4, e6140. <https://doi.org/10.1371/journal.pone.0006140>
- Western, D., Waithaka, J., & Kamanga, J. (2015). Finding space for wildlife beyond national parks and reducing conflict through community-based conservation: The Kenya experience. *PARKS*, 21, 51–62. <https://doi.org/10.2305/iucn.ch.2014.parks-21-1dw.en>
- Western, G. (2019). *Conflict or Coexistence: Human-lion relationships in Kenya's southern Maasailand and beyond*. Thesis.
- Wilson, E. O. (2016). *Half-earth: Our planet's fight for life*. New York, NY: WW Norton & Company.
- World Parks Congress (WPC). (2014). Summary of the International Union for Conservation of Nature (IUCN). Retrieved from <https://enb.iisd.org/iucn/wpc/2014/html/crsvol89num16e.html>
- WWF. (1980). *World conservation strategy: Living resource conservation for sustainable development*. Gland, Switzerland: IUCN.
- Yahdjian, L., Sala, O. E., & Havstad, K. M. (2015). Rangeland ecosystem services: Shifting focus from supply to reconciling supply and demand. *Frontiers in Ecology and the Environment*, 13, 44–51. <https://doi.org/10.1890/140156>

## SUPPORTING INFORMATION

Additional supporting information may be found online in the Supporting Information section.

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